

Study on mangrove associated estuarine waters of Northeastern Bay of Bengal reveals potential diatom indicators of dissolved inorganic compounds

Abhishek Mukherjee, Subhajit Das, Sabyasachi Chakraborty and Tarun Kumar De*

Department of Marine Science, Calcutta University, 35, B.C. Road, Kolkata-700 019, India. *Email: abmsws@gmail.com, tarunde@yahoo.co.in.

Abstract. An extensive study (2011 and 2012) was performed to narrow down certain diatom species with bioindicator potentials of highly dynamic and well mixed tropical estuary. The Hooghly Estuary in the Northeastern coast of Bay of Bengal was chosen as the study sites. The comparison between mangrove impoverished and mangrove dominated estuarine rivers revealed intriguing variation in the phytoplankton species composition in response to the dissolved nitrate, phosphate and silicate concentrations, thus forming the kernel of the study by focusing on certain species with greater abundances pointing to their better adaptability and responsiveness to biotic/abiotic parameters in a well mixed estuary. *Coscinodiscus hyalinus*, *Coscinodiscus radiatus*, *Lioloma elongata*, *Skeletonema costatum*, *Synedra crystallinum*, *Thalassionema frauenfeldtii* and *Thalassionema nitzschioides* have expressed better as bioindicators.

Keywords: Bioindicator, *Coscinodiscus*, *Skeletonema*, *Synedra*, *Thalassionema*, Hooghly Estuary.

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Introduction

Diatoms are microscopic, unicellular, eukaryotic algae that are always in full contact with their immediate ambient environment (Smol and Stomer, 2010) and react to the changes in their environment in the most sensitive way. The diatoms owing to their siliceous frustules are entombed in sediments for many years, a trait which has rendered them useful for studying and interpreting ecological conditions of bygone eras. The ubiquitous distribution of the diatoms, their high specific diversity and their siliceous frustules, all enable diatoms to function as sound environmental indicators (Stevenson and Pan, 1990). "It is estimated that diatoms account for 40% of the total marine production of carbon (Nelson et al., 1995). The chief inorganic compounds which govern the growth and

proliferation of the diatoms are nitrate, phosphate and silicate apart from carbon.

Nitrogen is considered to be one the most significant growth constraints for phytoplanktons. Around 70% of nitrogen assimilation is governed by the phytoplankton on a global scale and nitrate has been most widely studied as the chief nitrogen source of the phytoplankton. Its concentration ranges from undetectable to $50 \mu\text{mol.L}^{-1}$ in oceans and increase manifolds in the coastal waters as a result of riverine run-off and upwelling. Nitrate, in coastal waters, normally ranges within $500 \mu\text{mol.L}^{-1}$. Orthophosphate is the only form of phosphorus that autotrophs are able to assimilate. In contrary to the freshwater environment, in estuarine and marine environments phosphate plays the most important role of a growth limiting nutrient. More conclusive evidence for the effect of

phosphorous on growth rate has been reported using culture of certain algal species (Fuhs, 1971). The diatom cell walls are silicified and known as frustules. They generally consist of hydrated amorphous silica with empirical formula ($\text{Si}_n\text{O}_{2n-(nx/2)}\text{OH}_{nx}$), with $x \leq 4$. According to Tréguer et al. (1995), the use of silicon by diatoms dominates the biogeochemical cycling of Si in the sea, with each atom of Si weathered from land passing through a diatom on average 39 times before burial in the sea. Cell wall silicification and silicic acid transport are tightly coupled to the cell cycle, which results in a dependency in the extent of silicification on growth rate. According to Martin-Jézéquel et al. (2000) silica dissolution is an important part of diatom cellular silicon metabolism, because dissolution must be prevented in the living cell, and because much of the raw material for mineralization in natural assemblages is supplied by dissolution of dead cells.

Diatom indicators have been documented from time to time in aquatic environments such as lakes and lagoon but, hardly any substantial data exist on the use of particular diatom species as bioindicators of eutrophic conditions in well mixed tropical estuarine ecosystems. Both mangrove associated water bodies and rivers without apparent mangrove vegetation on their flanks harbor unique microalgal communities, dominated mostly by diatoms in estuarine ecosystems. Comparative accounts from their natural set ups and their responses towards various physiochemical conditions may be useful in revealing probable species with affinity towards a certain condition. The primary objective of the study was to delineate the response of phytoplanktons to changes in the inorganic nutrients so that few species can be demarcated with potential of being used as bioindicators to nutrient enriched environments in well mixed tropical estuaries.

Materials and methods

Study area

The sampling stations were Kachuberia (21° 51' 39" N, 88° 08' 37" E), Chemaguri (21° 40' 43" N, 88° 07' 28" E) and Gangasagar (21° 38' 00" N,

88° 05' 00" E) located in the Hooghly Estuary (Figure 1). The selection of these locales were mainly based on their clearly discernible salinity differences and mangrove vegetation covers, ranging from region with high riverine influence (Kachuberia) as mangrove impoverished zone, to brackish water regions and mangrove vegetation (Chemaguri) and marine dominated region (Gangasagar) with sparse mangrove cover. Three stations at Chemaguri were chosen to be sites with mangrove vegetation dominance (21° 40' 43" N and 88° 07' 28" E, 21° 40' 48" N and 88° 09' 04" E, and 21° 40' 57.54" N and 88° 09' 09.49" E). Results from mangrove dominated regions and mangrove impoverished regions have been grouped under the two headings and average data from the stations under respective groups has been used for the present manuscript. The Hooghly Estuary is the first deltaic offshoot of the River Ganges. It is a coastal plain estuary and lies approximately between 21° 32' N-22° 40' N and 88° 05' E-89° 00' E and the lower 80 km of the estuary flows through the western part of the marshy coastal areas covered with thick mangrove vegetation and is known worldwide as the Sundarbans.

Three primary sampling stations were selected due to their variation in salinity, ranging from river dominated (Kachuberia), brackish water (Chemaguri) and up to marine influenced (Ganga Sagar). The entire Hooghly Estuary was devastated by a severe tropical super cyclone and the study performed on it should help to find valuable information about the effect of such a catastrophe on microphytoplanktons population of a well mixed estuary (Hooghly Estuary is one of the most important part of the largest deltaic system in the World).

The northeast coast of the Bay of Bengal is highly irregular near the land-ocean boundary and is dissected by a large number of strewing rivulets, creeks and tributaries. The River Hooghly is the main artery of the River Ganges and is morphed into swampy marshy lands near its estuarine gateway to the Bay of Bengal, with severe fluvial deposits. The Hooghly Estuary exchanges flow with the Bay of Bengal near the Sagar Island and its tidal

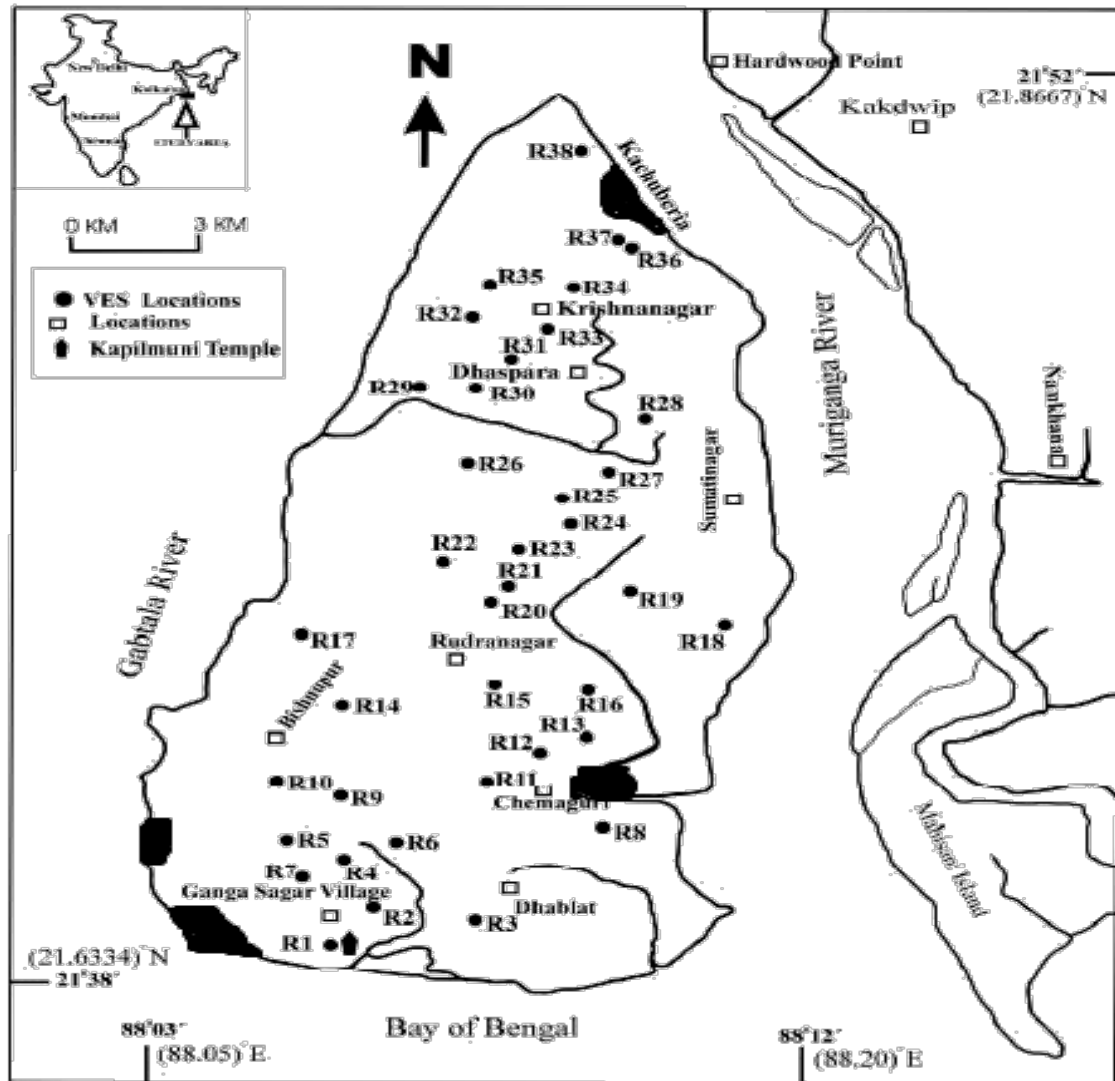


Figure 1. Map of the study sites with sampling stations (black blotches).

domain for nearly 250 km. the mouth of the estuary has a funnel shape and the predominant tidal regime is semidiurnal in nature. The vertical tide range at the mouth varies from 5.2 m during the spring tide to 1.8 m during the neap period. This is a well mixed estuary due to the intense tidal and wave actions with a meso-macrotidal setting (2.5-7 m amplitude). Mean current velocities are between 108 cm.s^{-1} and 117 cm.s^{-1} during high and low tides respectively. The climate of this region is characterized by the southwestern monsoon (July-October), northeast monsoon or post monsoon (November-February) and premonsoon (March-June); at least 70-80% of the rainfall occurs during the summer monsoon (SW monsoon), resulting in the high river discharge ($2,952\text{-}11,897 \text{ m}^3.\text{s}^{-1}$),

which gradually diminishes to $900\text{-}1500 \text{ m}^3.\text{s}^{-1}$ during non-monsoonal months (Biswas et al., 2004; Mukhopadhyay et al., 2006).

Data collection and sample analysis

Monthly data of surface water salinity, air temperature, water temperature, wind speed, dissolved oxygen, primary productivity and essential nutrients (nitrate, phosphate and silicate) were recorded and monthly phytoplanktonic sampling was performed from study sites for a total period of two years, 2011 and 2012. The essential meteorological back ground data of the days of the collections were procured from the usage of Shimadzu Weather Station. Salinity of the water was first

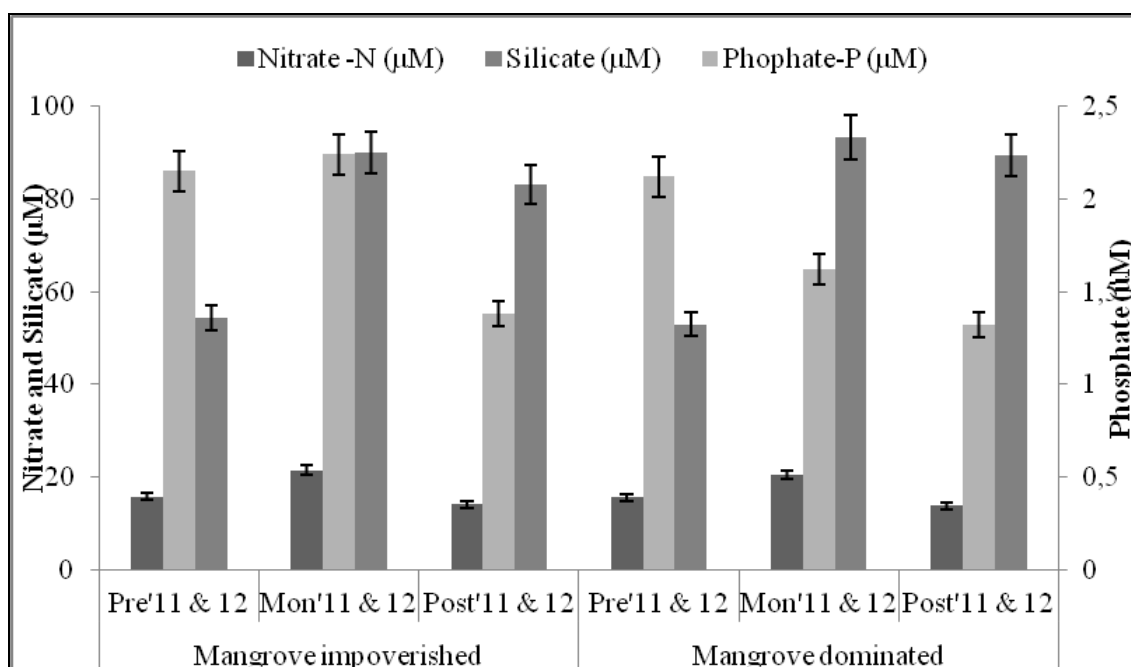


Figure 2. Seasonal fluctuations in the three chief nutrient parameters viz. nitrate-nitrogen, phosphate-phosphorus and silicate in mangrove impoverished and mangrove dominated waters of the Hooghly Estuary (error bars of the selected macronutrients with 5% value).

measured with the aid of refractometer and then estimated using argentometric method and the factor corrected values were put in.

Water sample for the estimation purposes was collected using the Niskin water sampler. For the estimation of phosphate, nitrate and silicate, the standard methods were followed (Grasshoff et al., 1983). Gross primary production and community respiration were measured *in situ* using the light and dark bottle method (Strickland and Parsons, 1972). The phytoplankton (diatom) sampling was performed using a handheld net [phytoplankton net (bolting silk no. 30, mesh size 20 µm) equipped with flow meter] from different stations in the Hooghly Estuary. The mesh size was fixed for the microphytoplanktons and not the nanoplanktons to maintain a certain degree of separation during the study and also due to the reason that increased turbidity rendered the usage of smaller mesh size impossible due to large quantities of silt clogging. The chief gear used in this purpose was country boats to ensure the least disturbance of the prevalent population of phytoplanktons. After the collection, the phytoplankton concentrates were transferred into 25 mL Tarson

polythene containers and 4% formalin along with lugol's iodine were used as preservatives depending on the need of clarification and specific preferences. The formalin is a very good agent for clearing cells off organic debris but it doesn't stain cells hence it is hard to discriminate between living, viable cells from dead, resuspended ones. The samples were then analyzed under light microscopes, bright field microscopes and phase contrast microscopes for their accurate and proper identification, using proper literature (effective and valid) (Desikachary, 1986-1989; Hasle and Syvertsen, 1997; Al-Kandari et al., 2009).

Another method for the quantification of phytoplanktons was used where 1 L water samples in triplicates were collected; each preserved with lugol's iodine solution and buffered formalin. After 72 h of sedimentation, the supernatant was filtered off (20 µm) and the phytoplanktons (diatoms, flagellates and dinoflagellates) were enumerated in a Sedgwick Rafter counting chamber. The data generated from both the methods were tallied and the deviation was negligible ensuring accurate enumeration.

Data compilation and statistical analyses

All data tabulation and compilation were accomplished using the MS-Excel software but for statistical analyses, the 'add-in' of Excel, the XLstat software was used and the principal component analysis tool of the data analysis package of the XLstat 12 was employed to deduce the Pearson Correlation Matrix.

Result

The study was performed with the objective of finding diatom species with bioindicator potentials in well mixed estuarine waters of Hooghly Estuary (Figure 1). Water samplings near mangrove mudflat (mangrove dominated) and water bodies with no apparent mangroves in sight (mangrove impoverished) were chosen due to their differences in physico-chemical parameters and may provide a wider platform for the diatom species to respond to, in terms of indicating a particular condition. The Figure 2 and the Table 1 summarize the comparative variations in water quality parameters such as nutrients (nitrate-nitrogen, phosphate-phosphorus and silicate), salinity, dissolved oxygen, primary, air and water temperatures, wind speed etc.

From Table 1, it is quite obvious that water quality parameters which mainly govern any phytoplankton (diatom) population assemblages in any aquatic environment are comparatively different on average in mangrove dominated and mangrove impoverished. Figure 2 depicts the variation in nutrients where it can be observed that among the three chief nutrients taken into consideration, dissolved phosphate actually followed a somewhat different pattern of seasonal fluctuation.

Tables 2, 3 and 4 comprehensively describes the seasonal comparative variations (in the relative abundances as well as the composition of the phytoplanktons) in the phytoplankton population (mainly diatoms) sampled from both mangrove dominated and mangrove impoverished water bodies of the Hooghly Estuary. The number of phytoplankton species in general was higher in mangrove dominated regions as compared to

mangrove impoverished sampling stations in all the three seasons. Again the relative abundances of phytoplankton species observed in the mangrove dominated region were found to be comparatively on the higher side as well.

The Table 5 depicts the species which had appeared in both the years of the study (2011 and 2012) in at least one of the three seasons, with few appearing in all the three. These phytoplanktons (diatoms) can best serve as indicators to the changes in their ambient environments. Species such as *Coscinodiscus radiatus*, *Coscinodiscus marginatus*, *Coscinodiscus lineatus*, *Ditylum brightwellii*, *Skeletonema costatum*, *Thalassionema frauenfeldtii*, *Thalassionema nitzschioides*, *Synedra crystallinum* etc are of more significance since they have been observed in both the years as well as all the three seasons, hence shall be able to reflect the effects of the seasonal changes in the selected physicochemical parameters on their composition and abundances and thus can be of greater use as bioindicators of well mixed tropical estuaries such as the one being dealt with presently. A very curious feature that can be observed from Table 5 is, almost all the species exhibited higher relative abundances in mangrove dominated waters than their abundances recorded from mangrove impoverished waters. Few species of *Coscinodiscus*, *Thalassionema* etc. however showed much higher abundances in mangrove impoverished waters.

A correlation was performed among relative abundance values of those species of phytoplanktons (diatoms) which were documented in both the years of the study, on all three seasons along with the chief nutrient parameters which is clearly exhibited in Table 6. From the table it is quite apparent that the correlation values were largely significant and a more common trend was that a species exhibited highly negative and positive correlations for a particular parameter in mangrove dominated and mangrove impoverished and vice versa. This suggests that although the parameter is similar, its exploitation might differ in mangrove dominated and mangrove impoverished conditions since the species composition differed leading to

Table 1. Average annual values of the background parameters with standard deviations (the data from mangrove dominated and impoverished waters were recorded seasonally and made average to capture the variations annually).

Parameters (2011 and 2012)	Mangrove dominated	Mangrove impoverished
PP (mgC m ⁻³ h ⁻¹)	30.20±5.45	29.77±5.49
Salinity (psu)	15.79±5.71	14.52±5.80
Air Temp. (aT°C)	26.50±3.05	26.52±3.67
Water Temp (wT°C)	27.45±3.87	27.51±3.34
Wind Speed (km.h ⁻¹)	6.66±1.52	6.25±1.82
D.O. (mgL ⁻¹)	5.82±1.49	5.93±1.40

Table 2. Comparative account of the premonsoonal (2011 and 2012) phytoplankton composition in mangrove impoverished and mangrove dominated waters of Hooghly Estuary in terms of relative abundance.

Mangrove Impoverished PRE'11 and 12		Mangrove dominated PRE'11 and 12	
Species	R.A. (%)	Species	R.A. (%)
<i>Asterionella japonicus</i>	0.9102	<i>Actinocyclus octonarius</i>	0.2249
<i>Aulacodiscus kittonii</i>	1.2136	<i>Asterionella japonicas</i>	2.142
<i>Bacteriastrum varians</i>	0.3034	<i>Bacillaria paxillifera</i>	0.5335
<i>Biddulphia (Odontella) longicuris</i>	2.8322	<i>Bacteriastrum varians</i>	1.4726
<i>Biddulphia (Odontella)mobiliensis</i>	1.2136	<i>Campylodiscus impressus</i>	0.3681
<i>Biddulphia (Odontella) regia</i>	1.011	<i>Ceratium cylindrus</i>	1.0389
<i>Biddulphia (Odontella) sinensis</i>	2.4272	<i>Ceratium furca</i>	1.6641
<i>Campylodiscus impressus</i>	0.6068	<i>Ceratium fusus</i>	1.8073
<i>Ceratium cylindrus</i>	0.9102	<i>Ceratium indicum</i>	0.7363
<i>Ceratium inflatum</i>	1.011	<i>Ceratium lineatum</i>	1.0014
<i>Chaetoceros balticum</i>	0.3034	<i>Ceratium trichoceros</i>	1.071
<i>Chaetoceros decipiens</i>	1.517	<i>Ceratium tripos</i>	3.0525
<i>Chaetoceros lorenzianus</i>	1.8204	<i>Chaetoceros compressus</i>	4.9923
<i>Chaetoceros quarctatus</i>	0.6068	<i>Chaetoceros constrictum</i>	1.0308
<i>Climacosphaenia monilifera</i>	1.2136	<i>Chaetoceros curvisetus</i>	4.1194
<i>Cocconeis scutellum</i>	0.405	<i>Chaetoceros danicus</i>	4.0297
<i>Coscinodiscus asteromphalus</i>	1.2136	<i>Chaetoceros decipiens</i>	4.0605
<i>Coscinodiscus centralis</i>	3.7417	<i>Chaetoceros externum</i>	1.6641
<i>Coscinodiscus concinnus</i>	4.6527	<i>Chaetoceros lorenzianus</i>	4.6817
<i>Coscinodiscus eccentricus</i>	0.6068	<i>Chaetoceros pseudocurvisetus</i>	0.7363
<i>Coscinodiscus gigas</i>	0.9102	<i>Chaetoceros peruvianus</i>	2.7498
<i>Coscinodiscus hyalinus</i>	4.551	<i>Chaetoceros quarctatus</i>	0.7363
<i>Coscinodiscus lineatus</i>	2.6297	<i>Corethron criophilum</i>	0.7363
<i>Coscinodiscus marginatus</i>	3.5399	<i>Coscinodiscus concinnus</i>	2.9493
<i>Coscinodiscus oculus-iridis</i>	1.2136	<i>Coscinodiscus eccentricus</i>	0.7363
<i>Coscinodiscus radiatus</i>	10.6191	<i>Coscinodiscus gigas</i>	0.2249
<i>Coscinodiscus tumidus</i>	0.3034	<i>Coscinodiscus hyalinus</i>	1.9184
<i>Cyclotella caspia</i>	2.3263	<i>Coscinodiscus marginatus</i>	0.3681
<i>Cyclotella striata</i>	1.2136	<i>Coscinodiscus radiates</i>	5.2132
<i>Cyclotella stylorum</i>	1.4161	<i>Cyclotella striata</i>	0.2249
<i>Ditylum brightwellii</i>	0.7076	<i>Ditylum brightwellii</i>	2.7766
<i>Gyrosigma angulatum</i>	0.6068	<i>Ditylum buchananii</i>	1.6065
<i>Haslea warwickea</i>	2.7306	<i>Hemidiscus kanayanus</i>	1.5543
<i>Hemidiscus cuneiformis</i>	1.011	<i>Lauderia annulata</i>	2.9748
<i>Hemidiscus kanayanus</i>	0.7084	<i>Leptocylindrus minimus</i>	2.1875
<i>Lioloma elongata</i>	1.4161	<i>Lioloma elongate</i>	0.3681
<i>Navicula transitans</i>	1.2136	<i>Okedenia inflexa</i>	0.7363

Table 2. Continued.

Mangrove impoverished PRE'11 and 12		Mangrove dominated PRE'11 and 12	
Species	R.A. (%)	Species	R.A. (%)
<i>Navicula</i> spp.	0.9102	<i>Noctiluca scintillans</i>	2.3388
<i>Nitzschia elongata</i>	1.2136	<i>Pleurosigma aestuarii</i>	1.1045
<i>Nitzschia sigma</i>	0.6068	<i>Proboscia alata</i> var. <i>indica</i>	0.5335
<i>Nitzschia sigmoidea</i>	2.7306	<i>Prorocentrum micans</i>	0.7363
<i>Nitzschia</i> spp.	0.3034	<i>Protoperidinium breviceps</i>	1.6493
<i>Noctiluca scintillans</i>	0.3034	<i>Protoperidinium depressum</i>	0.589
<i>Pinnularia viridis</i>	0.6068	<i>Protoperidinium divergens</i>	1.2959
<i>Pleurosigma aestuarii</i>	0.3034	<i>Protoperidinium pallidum</i>	1.2959
<i>Pleurosigma formosa</i>	0.3792	<i>Pseudonitzschia concatenatum</i>	1.071
<i>Pleurosigma</i> sp.	0.6068	<i>Pseudonitzschia delicatissima</i>	2.4781
<i>Pleurosigma</i> spp.	0.6068	<i>Pseudonitzschia seriata</i>	1.1567
<i>Proboscia alata</i> var. <i>indica</i>	0.7084	<i>Rhizosolenia hebetate</i>	1.8073
<i>Protoperidinium depressum</i>	0.6068	<i>Rhizosolenia setigera</i>	2.6253
<i>Protoperidinium divergens</i>	0.6068	<i>Trichodesmium erythraeum</i>	2.4245
<i>Protoperidinium pallidum</i>	0.6068	<i>Thalassionema frauenfeldtii</i>	5.2333
<i>Roperia tessellata</i>	0.6068	<i>Thalassionema nitzschioides</i>	5.3658
<i>Thalassionema frauenfeldtii</i>	9.608	<i>Skeletonema costatum</i>	6.6749
<i>Thalassionema nitzschioides</i>	6.6749	<i>Synedra crystallinum</i>	0.9102
<i>Thalassiosira condensata</i>	1.2136	<i>Synedra ulna</i>	0.3681
<i>Thalassiothrix longissima</i>	1.011		
<i>Skeletonema costatum</i>	0.405		
<i>Synedra crystallinum</i>	0.9102		
<i>Synedra ulna</i> var. <i>oxyrhyncus</i>	2.8315		

Table 3. Comparative account of the monsoonal (2011 and 2012) phytoplankton composition in mangrove impoverished and mangrove dominated waters of Hooghly Estuary in terms of relative abundance.

Mangrove impoverished MON'11 and 12		Mangrove dominated MON'11 and 12	
Species	R.A. (%)	Species	R.A. (%)
<i>Actinocyclus octonarius</i>	3.5014	<i>Amphora ostrearia</i>	0.3368
<i>Amphora ostrearia</i>	1.4005	<i>Amphora ovate</i>	0.3062
<i>Asterionella japonicus</i>	2.8011	<i>Aulacodiscus kittonii</i>	0.6329
<i>Aulacodiscus kittonii</i>	1.1204	<i>Bellerochea malleus</i>	0.689
<i>Bacteriastrum varians</i>	1.4005	<i>Biddulphia (Odontella) longicuris</i>	0.7758
<i>Biddulphia (Odontella) mobiliensis</i>	1.8674	<i>Biddulphia (Odontella) mobiliensis</i>	0.6125
<i>Biddulphia (Odontella) regia</i>	1.4005	<i>Biddulphia (Odontella) regia</i>	0.8452
<i>Biddulphia (Odontella) sinensis</i>	1.5406	<i>Biddulphia (Odontella) sinensis</i>	2.1029
<i>Ceratium inflatum</i>	1.2605	<i>Campylodiscus impressus</i>	0.6125
<i>Climacosphaenia</i> sp.	1.4005	<i>Ceratium furca</i>	1.4677
<i>Cocconeis pediculus</i>	1.4005	<i>Ceratium inflatum</i>	0.8575
<i>Chaetoceros curvisetum</i>	1.4005	<i>Ceratium tripos</i>	1.4677
<i>Chaetoceros lorenzianus</i>	0.7002	<i>Chaetoceros curvisetum</i>	1.4677
<i>Chaetoceros peruvianus</i>	0.7002	<i>Chaetoceros compressum</i>	1.4677
<i>Chaetoceros subtilis</i>	1.4005	<i>Chaetoceros decipiens</i>	1.4677
<i>Coscinodiscus asteromphalus</i>	2.661	<i>Chaetoceros pseudocurvisetus</i>	0.9187
<i>Coscinodiscus centralis</i>	1.2605	<i>Chaetoceros quarctatus</i>	0.6125
<i>Coscinodiscus concinnus</i>	1.3071	<i>Coscinodiscus centralis</i>	1.4904
<i>Coscinodiscus eccentricus</i>	1.1671	<i>Coscinodiscus concinnus</i>	0.8268
<i>Coscinodiscus granii</i>	1.6339	<i>Coscinodiscus granii</i>	1.8375
<i>Coscinodiscus hyalinus</i>	2.2408	<i>Coscinodiscus hyalinus</i>	1.6966

Table 3. Continued.

Mangrove impoverished MON'11 and 12		Mangrove dominated MON'11 and 12	
Species	R.A. (%)	Species	R.A. (%)
<i>Coscinodiscus lineatus</i>	1.6806	<i>Coscinodiscus lineatus</i>	1.4547
<i>Coscinodiscus marginatus</i>	2.8011	<i>Coscinodiscus marginatus</i>	1.2403
<i>Coscinodiscus radiatus</i>	2.1008	<i>Coscinodiscus radiatus</i>	2.2908
<i>Cyclotella striata</i>	1.4005	<i>Cyclotella caspia</i>	3.6751
<i>Diatoma anceps</i>	1.4005	<i>Cyclotella striata</i>	0.6125
<i>Ditylum brightwellii</i>	3.3613	<i>Diploneis sp.</i>	0.4593
<i>Ditylum buehananii</i>	1.4005	<i>Ditylum brightwellii</i>	0.2848
<i>Hemidiscus kanayanus</i>	1.1204	<i>Gyrosigma angulatum</i>	0.3675
<i>Lioloma elongata</i>	10.6442	<i>Gyrosigma scalproides</i>	0.6125
<i>Leptocylindrus danicus</i>	1.4005	<i>Haslea warwickea</i>	0.245
<i>Navicula pupalis</i>	1.1204	<i>Lioloma elongata</i>	0.8575
<i>Navicula sp.</i>	1.1204	<i>Melosira nummuloidea</i>	1.6384
<i>Navicula pediculus</i>	1.4005	<i>Microcystis robusta</i>	4.5938
<i>Navicula transitans</i>	1.4005	<i>Navicula balticum</i>	0.6125
<i>Nitzschia sigma</i>	0.7002	<i>Navicula transitans</i>	0.4593
<i>Nitzschia sigmoidea</i>	1.1204	<i>Navicula sp.</i>	0.3062
<i>Pleurosigma elongatum</i>	2.1008	<i>Navicula pupalis</i>	0.9187
<i>Proboscia alata var. indica</i>	1.4005	<i>Nitzschia bilobata</i>	0.6125
<i>Pseudonitzschia seriata</i>	1.8674	<i>Nitzschia elongata</i>	0.7758
<i>Rhizosolenia setigera</i>	2.1008	<i>Nitzschia sigma</i>	0.5512
<i>Skeletonema costatum</i>	5.042	<i>Nitzschia sigmoidea</i>	0.8039
<i>Skeletonema grevillei</i>	4.4817	<i>Noctiluca scintillans</i>	0.8268
<i>Surirella ovata</i>	1.4005	<i>Pinnularia viridis</i>	0.98
<i>Synedra crystallinum</i>	2.1708	<i>Pleurosigma sp.</i>	0.6125
<i>Synedra ulna</i>	1.2605	<i>Pleurosigma elongatum</i>	1.8375
<i>Synedra ulna var. oxyrhynchus</i>	2.1008	<i>Prorocentrum micans</i>	0.6125
<i>Thalassionema nitzschioides</i>	2.4509	<i>Rhizosolenia setigera</i>	0.6125
<i>Thalassionema frauenfeldtii</i>	1.4007	<i>Skeletonema costatum</i>	15.6804
<i>Thalassiosira condensata</i>	1.6806	<i>Skeletonema grevillei</i>	7.5952
<i>Thalassiothrix longissima</i>	1.4007	<i>Surirella ovata</i>	0.3062
<i>Trichodesmium erythraeum</i>	1.4007	<i>Synedra crystallinum</i>	3.2157
		<i>Thalassiosira condensata</i>	5.2926
		<i>Thalassionema nitzschioides</i>	8.7057
		<i>Thalassionema frauenfeldtii</i>	0.9968
		<i>Triceratium favus</i>	0.58
		<i>Trichodesmium erythraeum</i>	14.2055

Table 4. Comparative account of the post monsoonal (2011 and 2012) phytoplankton composition in mangrove impoverished and mangrove dominated waters of Hooghly estuary in terms of relative abundance.

Mangrove impoverished POST'11 and 12		Mangrove dominated POST'11 and 12	
Species	R.A. (%)	Species	R.A. (%)
<i>Actinocyclus senarius</i>	1.1985	<i>Amphora ovate</i>	0.2921
<i>Bellerochea malleus</i>	1.1985	<i>Aulacodiscus kittonii</i>	0.4381
<i>Biddulphia (Odontella) regia</i>	2.0599	<i>Bellerochea malleus</i>	1.0954
<i>Biddulphia (Odontella) mobiliensis</i>	0.5617	<i>Biddulphia (Odontella) longicruris</i>	0.5969
<i>Biddulphia (Odontella) sinensis</i>	2.2097	<i>Biddulphia (Odontella) mobiliensis</i>	0.5477
<i>Cerataulina pelagica</i>	2.0224	<i>Biddulphia (Odontella) regia</i>	0.7795
<i>Ceratium cylindrus</i>	1.1235	<i>Biddulphia (Odontella) sinensis</i>	1.7066

Table 4. Continued.

Mangrove impoverished	POST'11 and 12	Mangrove dominated	POST'11 and 12
Species	R.A. (%)	Species	R.A. (%)
<i>Ceratium indicum</i>	1.1235	<i>Ceratium cylindrus</i>	1.3144
<i>Chaetoceros compressus</i>	2.0599	<i>Ceratium furca</i>	0.8069
<i>Chaetoceros decipiens</i>	1.6853	<i>Ceratium indicum</i>	1.0954
<i>Chaetoceros lorenzianus</i>	2.7715	<i>Ceratium inflatum</i>	0.5887
<i>Chaetoceros peruvianus</i>	2.9213	<i>Ceratium tripos</i>	1.0954
<i>Chaetoceros quarctatus</i>	1.2359	<i>Chaetoceros compressus</i>	0.8215
<i>Coscinodiscus asteromphalus</i>	1.7977	<i>Chaetoceros curvisetus</i>	1.6431
<i>Coscinodiscus centralis</i>	5.131	<i>Chaetoceros danicus</i>	0.3833
<i>Coscinodiscus gigas</i>	1.7602	<i>Chaetoceros decipiens</i>	0.3067
<i>Coscinodiscus granii</i>	1.6479	<i>Chaetoceros lorenzianus</i>	0.6627
<i>Coscinodiscus hyalinus</i>	4.4569	<i>Chaetoceros pseudocurvisetus</i>	1.0954
<i>Coscinodiscus lineatus</i>	2.322	<i>Chaetoceros quarctatus</i>	0.3505
<i>Coscinodiscus marginatus</i>	3.7453	<i>Coscinodiscus asteromphalus</i>	1.2323
<i>Coscinodiscus radiatus</i>	5.0936	<i>Coscinodiscus centralis</i>	0.6755
<i>Cyclotella caspia</i>	1.7977	<i>Coscinodiscus concinnus</i>	1.7855
<i>Cyclotella striata</i>	0.5617	<i>Coscinodiscus granii</i>	1.2432
<i>Cyclotella stylorum</i>	1.1235	<i>Coscinodiscus hyalinus</i>	1.089
<i>Diploneis sp.</i>	0.5617	<i>Coscinodiscus lineatus</i>	1.2323
<i>Ditylum buchananii</i>	2.0599	<i>Coscinodiscus marginatus</i>	0.931
<i>Ditylum brightwellii</i>	8.4269	<i>Coscinodiscus radiatus</i>	6.7915
<i>Haslea warwickea</i>	2.9213	<i>Diploneis sp.</i>	0.5477
<i>Hemidiscus kanayanus</i>	0.8239	<i>Ditylum brightwellii</i>	1.4459
<i>Lauderia annulata</i>	0.8239	<i>Ditylum buchananii</i>	0.3067
<i>Lioloma elongata</i>	0.8239	<i>Gyrosigma angulatum</i>	0.4929
<i>Navicula transitans</i>	2.397	<i>Guinardia sp.</i>	0.23
<i>Navicula angulatum</i>	1.7977	<i>Haslea warwickea</i>	0.3286
<i>Nitzschia bilobata</i>	3.7078	<i>Lauderia sp.</i>	0.23
<i>Nitzschia sigma</i>	1.6853	<i>Lioloma elongata</i>	0.8259
<i>Nitzschia sigmaidea</i>	0.8239	<i>Melosira nummuloidea</i>	5.9152
<i>Noctiluca scintillans</i>	1.9475	<i>Microcystis robusta</i>	8.2155
<i>Okedenia inflexa</i>	1.1235	<i>Navicula sp.</i>	1.0954
<i>Pinnularia rectangulata</i>	2.4719	<i>Navicula balticum</i>	0.5477
<i>Porosira sp.</i>	1.1985	<i>Navicula pupalis</i>	0.2738
<i>Pseudonitzschia seriata</i>	1.1985	<i>Navicula transitans</i>	0.4107
<i>Rhizosolenia styliformis</i>	0.8239	<i>Nitzschia bilobata</i>	0.5477
<i>Rhizosolenia setigera</i>	1.6853	<i>Nitzschia sigmaidea</i>	0.6586
<i>Skeletonema costatum</i>	1.6853	<i>Noctiluca scintillans</i>	0.7394
<i>Skeletonema grevillei</i>	1.1235	<i>Pinnularia rectangulata</i>	1.0954
<i>Synedra crystallinum</i>	2.9213	<i>Pinnularia viridis</i>	0.6846
<i>Thalassionema frauenfeldtii</i>	1.2359	<i>Pleurosigma sp.</i>	1.6431
<i>Thalassiosira condensata</i>	1.7977	<i>Pleurosigma aestuarii</i>	1.1227
<i>Thalassionema nitzschioides</i>	1.1985	<i>Pleurosigma elongata</i>	1.0954
<i>Trichodesmium erythraeum</i>	1.1985	<i>Pleurosigma rectangulata</i>	0.23
		<i>Prorocentrum micans</i>	0.5477
		<i>Rhizosolenia styliformis</i>	0.3833
		<i>Skeletonema costatum</i>	17.9647
		<i>Skeletonema grevillei</i>	6.0247
		<i>Synedra ulna</i>	0.3286
		<i>Synedra crystallinum</i>	2.8754
		<i>Thalassionema frauenfeldtii</i>	0.9968
		<i>Thalassionema nitzschioides</i>	3.1931
		<i>Thalassiosira condensata</i>	4.0676
		<i>Triceratium favus</i>	0.2738
		<i>Trichodesmium erythraeum</i>	0.7302

Table 5. Spices of diatoms which occurred in both the years of the study tenure, in mangrove impoverished and mangrove dominated estuarine waters, serving as tools to understand the effect of the two different ambiances on the existing phytoplankton populations. The bolder names are of those which have appeared in more than one season.

Pre'11 and 12	M.I. R.A. (%)	M.D. R.A. (%)	Mon'11 and 12	M.I. R.A. (%)	M.D. R.A. (%)	Post '11 and 12	M.I. R.A. (%)	M.D. R.A. (%)
<i>Asterionella japonicus</i>	0.9102	2.142	<i>Amphora ostrearia</i>	1.4005	0.3368	<i>Bellerochea malleus</i>	1.1985	1.0954
<i>Bacteriastrum varians</i>	0.3034	1.4726	<i>Aulacodiscus kittonii</i>	1.1204	0.6329	<i>Biddulphia (Odontella) regia</i>	2.0599	0.5477
<i>Campylodiscus impressus</i>	0.6068	0.3681	<i>Biddulphia (Odontella) mobiliensis</i>	1.8674	0.6125	<i>Biddulphia (Odontella) mobiliensis</i>	0.5617	0.7795
<i>Ceratium cylindrus</i>	0.9102	1.0389	<i>Biddulphia (Odontella) regia</i>	1.4005	0.8452	<i>Biddulphia (Odontella) sinensis</i>	2.2097	1.7066
<i>Chaetoceros decipiens</i>	1.517	4.0605	<i>Biddulphia (Odontella) sinensis</i>	1.5406	2.1029	<i>Ceratium cylindrus</i>	1.1235	1.3144
<i>Chaetoceros lorenzianus</i>	1.8204	4.6817	<i>Ceratium inflatum</i>	1.2605	0.8575	<i>Ceratium indicum</i>	1.1235	1.0954
<i>Chaetoceros quarctatus</i>	0.6068	0.7363	<i>Chaetoceros curvisetum</i>	1.4005	1.4677	<i>Chaetoceros compressus</i>	2.0599	0.8215
<i>Coscinodiscus concinnus</i>	4.6527	2.9493	<i>Coscinodiscus centralis</i>	1.2605	1.4904	<i>Chaetoceros decipiens</i>	1.6853	0.3067
<i>Coscinodiscus eccentricus</i>	0.6068	0.7363	<i>Coscinodiscus concinnus</i>	1.3071	0.8268	<i>Chaetoceros lorenzianus</i>	2.7715	0.6627
<i>Coscinodiscus gigas</i>	0.9102	0.2249	<i>Coscinodiscus granii</i>	1.6339	1.8375	<i>Chaetoceros quarctatus</i>	1.2359	0.3505
<i>Coscinodiscus hyalinus</i>	4.551	1.9184	<i>Coscinodiscus hyalinus</i>	2.2408	1.6966	<i>Coscinodiscus asteromphalus</i>	1.7977	1.2323
<i>Coscinodiscus marginatus</i>	3.5399	0.3681	<i>Coscinodiscus lineatus</i>	1.6806	1.4547	<i>Coscinodiscus centralis</i>	5.131	0.6755
<i>Coscinodiscus radiatus</i>	10.6191	5.2132	<i>Coscinodiscus marginatus</i>	2.8011	1.2403	<i>Coscinodiscus granii</i>	1.6479	1.2432
<i>Cyclotella striata</i>	1.2136	0.2249	<i>Coscinodiscus radiatus</i>	2.1008	2.2908	<i>Coscinodiscus hyalinus</i>	4.4569	1.089
<i>Ditylum brightwellii</i>	0.7076	2.7766	<i>Cyclotella striata</i>	1.4005	0.6125	<i>Coscinodiscus lineatus</i>	2.322	1.2323
<i>Hemidiscus kanayanus</i>	0.7084	1.5543	<i>Ditylum brightwellii</i>	3.3613	0.2848	<i>Coscinodiscus marginatus</i>	3.7453	0.931
<i>Lioloma elongata</i>	1.4161	0.3681	<i>Lioloma elongata</i>	10.6442	0.8575	<i>Coscinodiscus radiatus</i>	5.0936	6.7915
<i>Noctiluca scintillans</i>	0.3034	2.3388	<i>Navicula transitans</i>	1.4005	0.4593	<i>Diploneis</i> sp.	0.5617	0.5477
<i>Pleurosigma aestuarii</i>	0.3034	1.1045	<i>Nitzschia sigma</i>	0.7002	0.5512	<i>Ditylum buchananii</i>	2.0599	1.4459
<i>Proboscia alata</i> var. <i>indica</i>	0.7084	0.5335	<i>Nitzschia sigmoidea</i>	1.1204	0.8039	<i>Ditylum brightwellii</i>	8.4269	0.3067
<i>Protoperidinium depressum</i>	0.6068	0.589	<i>Pleurosigma elongatum</i>	2.1008	1.8375	<i>Haslea warwickea</i>	2.9213	0.3286
<i>Protoperidinium divergens</i>	0.6068	1.2959	<i>Rhizosolenia setigera</i>	2.1008	0.6125	<i>Lauderia annulata</i>	0.8239	0.23
<i>Protoperidinium pallidum</i>	0.6068	1.2959	<i>Skeletonema costatum</i>	5.042	15.6804	<i>Lioloma elongate</i>	0.8239	0.8259
<i>Thalassionema frauenfeldtii</i>	9.608	5.2333	<i>Skeletonema grevillei</i>	4.4817	7.5952	<i>Navicula transitans</i>	2.397	0.4107
<i>Thalassionema nitzschioides</i>	6.6749	5.3658	<i>Surirella ovata</i>	1.4005	0.3062	<i>Nitzschia sigmoidea</i>	0.8239	0.6586
<i>Skeletonema costatum</i>	0.405	6.6749	<i>Synedra crystallinum</i>	2.1708	3.2157	<i>Noctiluca scintillans</i>	1.9475	0.7394
<i>Synedra crystallinum</i>	0.9102	0.9102	<i>Thalassionema nitzschioides</i>	2.4509	5.2926	<i>Pinnularia rectangulata</i>	2.4719	1.0954
<i>Synedra ulna</i> var. <i>oxyrhynchus</i>	2.8315	0.3681	<i>Thalassionema frauenfeldtii</i>	1.4007	8.7057	<i>Rhizosolenia styliformis</i>	0.8239	0.3833
			<i>Thalassiosira condensata</i>	1.6806	0.9968	<i>Skeletonema costatum</i>	1.6853	17.964 7
			<i>Trichodesmium erythraeum</i>	1.4007	14.2055	<i>Skeletonema grevillei</i>	1.1235	6.0247

Table 5. Continued.

Pre'11 and 12	M.I. R.A. (%)	M.D. R.A. (%)	Mon'11 and 12	M.I. R.A. (%)	M.D. R.A. (%)	Post '11 and 12	M.I. R.A. (%)	M.D. R.A. (%)
						<i>Synedra crystallinum</i>	2.9213	2.8754
						<i>Thalassionema frauenfeldtii</i>	1.2359	0.9968
						<i>Thalassiosira condensata</i>	1.7977	3.1931
						<i>Thalassionema nitzschioides</i>	1.1985	4.0676
						<i>Trichodesmium erythraeum</i>	1.1985	0.7302

M.I. = Mangrove impoverished; M.D. = Mangrove dominated.

Table 6. The following table is depicting the major diatoms species and their responses to various nutrient fluctuations in mangrove dominated and impoverished waters of Hooghly Estuary, through the correlation with the nutrients contents and relative abundances.

	M.I.	M.D.	M.I.	M.D.	M.I.	M.D.
Species	Nitrate-N	Nitrate-N	Phosphate-P	Phosphate-P	Silicate	Silicate
<i>Coscinodiscus hyalinus</i>	-0.9681	+0.5115	-0.5504	+0.9188	-0.6737	-0.6408
<i>Coscinodiscus marginatus</i>	-0.9999	+0.8266	+0.7419	-0.4505	+0.4672	+0.8057
<i>Coscinodiscus radiatus</i>	-0.6132	-0.9969	+0.0745	-0.2079	-0.9849	-0.2571
<i>Ditylum brightwellii</i>	-0.3852	-0.2538	-0.9045	+0.9257	+0.6358	-0.9966
<i>Lioloma elongata</i>	+0.9866	+0.3019	+0.6230	-0.9056	+0.6048	+0.9995
<i>Synedra crystallinum</i>	-0.0715	+0.3766	-0.7218	-0.8691	+0.8483	+0.9988
<i>Skeletonema costatum</i>	+0.8834	+0.0564	+0.3413	-0.9824	+0.8270	+0.9605
<i>Thalassionema frauenfeldtii</i>	-0.2853	+0.9534	+0.4308	+0.4259	-0.9806	+0.0321
<i>Thalassionema nitzschioides</i>	-0.0864	+0.6801	+0.6037	+0.8157	-0.9211	-0.4657

M.I. = Mangrove impoverished; M.D. = Mangrove dominated.

different interspecific competition stress. All the diatom species have expressed significantly high correlation values with nutrients but very few have actually reflected a constant trend to be utilized as indicators to a certain conditions. But the ones that have expressed such affinities can truly serve as potential bioindicators of well mixed estuarine aquatic ecosystems.

Discussion

The water bodies adjacent to mangrove mudflats and those that are devoid of any mangrove mudflats were chosen due to their variations in the selected background physicochemical parameters which are clearly depicted in Figure 2 and Table 1. Although fluctuation in seasonal dissolved nitrate-nitrogen and dissolved silicate contents in mangrove impoverished and mangrove dominated

well mixed estuarine waters were found to be more or less similar, the difference in phosphate concentration was quite obvious during the tenure of the study. The reason behind dissolved phosphate contents in mangrove dominated regions to be high is due to the fact that mangrove mudflats harbor large number of phosphate solubilizing bacteria (De et al., 2012) or PSB and they through the formation of low molecular weight organic acid, bind and release the inorganic phosphate into solution phase. Hence, higher number of PSB means greater solubilization of phosphate.

Primary productivity in the mangrove dominated waters appeared to be comparatively higher than the mangrove impoverished waters (Table 1), a fact that can be vindicated from the Tables 2, 3 and 4 which depict comparatively higher number of phytoplankton species (diatoms

mainly) in mangrove dominated waters. Greater dissolved phosphate might lead to conditions conducive enough to support a higher population density.

Dissolved oxygen however exhibited a more or less similar fluctuation in both the specific study sites with slightly lower values recorded from mangrove dominated estuaries.

Salinity was also found to be higher in mangrove dominated waters since the mangrove dominated sampling stations were situated closer to the estuary mouth, compared to the mangrove impoverished stations that included stations with riverine dominance along with marine influence, eventually decreasing the average salinity of the surface water. Another fact that can contribute to the surface water salinity is the prevailing wind speed, which was recorded to be relatively high over mangrove dominated sampling sites than their mangrove impoverished counterparts. This could have triggered higher rates of evaporation resulting in higher salinity of the surface water.

From the Table 2, 3 and 4, the fact that phytoplankton population density in mangrove dominated waters were much higher compared to the mangrove impoverished waters. The above mentioned governing physicochemical parameters can explain phytoplankton proliferation in mangrove dominated water. The three tables are depicting seasonal variations in the phytoplankton (diatom mainly) relative abundances. Table 5 describes the phytoplankton species which have been observed in mangrove dominated and mangrove impoverished waters in a given season. From this table, it can be observed that in general the species appearing in mangrove dominated waters had higher relative abundances than when they appeared in mangrove impoverished waters with exceptions like *Coscinodiscus* spp, *Odontella* spp, *Thalassionema frauenfeldtii* and *Thalassionema nitzschioides* etc. These species showed very high relative abundances in mangrove impoverished waters. It has to be borne in mind that the species were not collected at different stations on different days, but on the same days of collection, hence emphasizing more on the growth conducive factors of the

mangrove dominated waters, viz. dissolved phosphate.

Among the species mentioned in Table 5, only a handful of phytoplankton species were observed to occur in all the three seasons, both the years of study as well as both in mangrove dominated and in mangrove impoverished estuarine waters.

A correlation study was performed among these species in terms of their relative abundances in both mangrove dominated and mangrove impoverished waters and the chief governing nutrients (Table 6). From the table it is quite obvious that almost all the species which have appeared in mangrove dominated and impoverished waters throughout the year have generated significant correlation values, both positive and negative. Mostly a particular parameter has shown positive and negative values for the same parameter considered in mangrove dominated and impoverished waters. This could be due to the fact that both the different study zones might have had different constraints and various species composition in a given zone rendered the exploitation of a particular parameter specific to that zone only, depending on background parameters, interspecific competition stress of a zone as well as due to factors unaccounted for.

Only those with a consistent correlation value have been chosen to serve as potential indicators to nutrient contents in well mixed estuarine waters. From table 6, the following species viz., *Coscinodiscus hyalinus* [M.I. silicate: -0.6737 and M.D. silicate: -0.6408], *Coscinodiscus radiatus* [M.I. nitrate: -0.6132 and M.D. nitrate: -0.9939; M.I. silicate: -0.9849 and M.D. silicate: -0.2571], *Lioloma elongata* [M.I. nitrate: +0.9866 and M.D. nitrate: +0.3019; M.I. silicate: +0.6048 and M.D. silicate: +0.9995], *Skeletonema costatum* [M.I. silicate: +0.8270 and M.D. silicate: +0.9605], *Synedra crystallinum* [M.I. phosphate: -0.7218 and M.D. phosphate: -0.8691; M.I. silicate: +0.8483 and M.D. silicate: +0.9988], *Thalassionema frauenfeldtii* [M.I. phosphate: +0.4308 and M.D. phosphate: +0.4259; M.I. silicate: -0.9806 and M.D. silicate: +0.0321] and *Thalassionema nitzschioides* [M.I. phosphate: +0.6037 and M.D. phosphate: +0.8157; M.I. silicate: -0.9277 and M.D.

silicate: -0.4657], where M.I. = Mangrove Impoverished, M.D. = Mangrove Dominated.

From the above stated data set it is quite obvious that these few diatom species do possess the potential of being established as probable bioindicators of various nutrients in well mixed tropical estuaries such as the Hooghly Estuary. The positive correlation implies that the relative abundance of a particular species increased with increase of a specific nutrient content and inversely negative correlation means the relative abundance is lowest when the nutrient concentration is highest, which could be due to over proliferation of a competitive species exploiting the increased nutrient or it could be because the species in question has utilized the nutrient to increase in number, rendering the dissolved nutrient low in concentration. Both the scenarios are highly probable in any well mixed estuarine ecosystem.

Conclusion

Although the study encompassed a tenure of two years (2011 and 2012), but the results generated from it is of immense significance since it was aimed at establishing probable bioindicator phytoplankton species (mainly diatoms due to their preponderance) in a well mixed estuary. An estuarine river system might be influenced by the presence and absence of mangrove forests along its bank, a fact prominently shown in the present study. Any changes in the ambient water quality parameters governing phytoplankton species composition and their relative abundances in a population gets reflected through the absence, presence or proliferation of certain species. The diatoms are the most dominant phytoplanktons in a well mixed estuarine aquatic ecosystem and those species which expressed a very strong degree of affinity towards a certain trophic condition have been documented through rigorous analyses of samples from natural set ups, thereby ensuring the highest possible authenticity regarding the applicability and implication of the study results. Establishment of phytoplanktons with potential as bioindicators in well mixed estuarine waters will ensure better

monitoring of the ever changing and highly dynamic ecosystem in a holistic way, along with the existing abiotic estimation and survey methods.

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Conflict of interest statement

Authors declare that they have no conflict of interests.

References

- Al-Kandari, M.; Al-Yamani, F. Y.; Al-Rifaie, K. **Marine phytoplankton atlas of Kuwait's waters**. Safat: Kuwait Institute for Scientific Research, 2009.
- Biswas, H.; Mukhopadhyay, S. K.; De, T. K.; Sen, S.; Jana, T. K. Biogenic controls on the air-water carbon dioxide exchange in the Sundarban mangrove environment, Northeast coast of Bay of Bengal, India. **Limnol. Oceanogr.**, v. 49, n. 1, p. 95-101, 2004.
- De, T. K.; Sarkar, T. K.; Mukherjee, A.; Maiti, T. K.; Das, S. Abundance and occurrence of phosphate solubilizing bacteria and phosphatase in sediment of Hooghly Estuary, North East Coast of Bay of Bengal, India. **Journal of Coastal Development**, v. 15, n. 1, p. 9-16, 2011.
- Desikachary, T. V. **Atlas of diatoms**. Madras: Madras Science Foundation, 1986-1989.
- Fuhs, G. W.; Demmerle, S. D.; Connelli, E.; Chen, M. Characterization of phosphorus limited plankton algae. **Environ. Conserv. Tech. Pap.**, v. 6, p. 1-40, 1971.
- Grasshoff, K.; Erhardt, M.; Kremling, K. **Methods of seawater analysis**. Weinheim: Verlag Chemie, 1983.
- Hasle, G. R.; Syvertsen, E. R. Marine diatoms. In: Tomas, C. R. (Ed). **Identifying marine phytoplankton**. London: Academic Press, 1994. p. 5-385.

- Martin-Jézéquel, V.; Hildebrand, M.; Brzezinski, M.A. Silicon metabolism in diatoms: implications for growth. **J. Phycol.**, v. 36, p. 821-840, 2000.
- Mukhopadhyay, S. K.; Biswas, H.; De, T. K.; Jana, T. K. Fluxes of nutrients from the tropical River Hoogly at the land-ocean boundary of Sundarban, NE coast of Bay of Bengal, India. **J. Mar. Sys.**, v. 62, p. 9-21, 2006.
- Nelson, D. M.; Tréguer, P.; Brzezinski, M. A.; Leynert, A.; Quéguiner, B. Production and dissolution of biogenic silica in the ocean: revised global estimates, comparison with regional data and relationship to biogenic sedimentation. **Global Biogeochemical Cycle**, v. 9, n. 3, p. 359-372, 1995.
- Smol, J. P.; Stroemer, E. F. **The diatoms: applications for the environmental and Earth Sciences**. 2 ed. Cambridge: Cambridge University Press, 2010.
- Stevenson, R. J.; Pan, Y. Diatoms as indicators of coastal palaeo-environments and relative sea-level change. In: Stroemer, E. F.; Smol, J. P. (ed.). **The diatoms: applications for the environmental and earth sciences**. New York: Cambridge University Press, 1999. p. 277-418.
- Strickland, J. D. H.; Parsons, T. R. **A practical handbook of seawater analysis**. 2 ed. Canada: Fish Res. Board, 1972.
- Tréguer, P.; Nelson, D. M.; Van Bennekom, A. J.; DeMaster, D. J.; Leynert, A.; Quéguiner, B. The silica balance in the world ocean: a reestimate. **Science**, v. 268, p. 375-379, 1995.